





Long-term differences in extinction risk among the seven forms of rarity

Paul G. Harnik^{1,*,†}, Carl Simpson² and Jonathan L. Payne¹

¹Department of Geological and Environmental Sciences, Stanford University, Stanford, CA 94305, USA ²Museum für Naturkunde, Leibniz Institute for Research on Evolution and Biodiversity at the Humboldt University Berlin, Berlin 10115, Germany

Rarity is widely used to predict the vulnerability of species to extinction. Species can be rare in markedly different ways, but the relative impacts of these different forms of rarity on extinction risk are poorly known and cannot be determined through observations of species that are not yet extinct. The fossil record provides a valuable archive with which we can directly determine which aspects of rarity lead to the greatest risk. Previous palaeontological analyses confirm that rarity is associated with extinction risk, but the relative contributions of different types of rarity to extinction risk remain unknown because their impacts have never been examined simultaneously. Here, we analyse a global database of fossil marine animals spanning the past 500 million years, examining differential extinction with respect to multiple rarity types within each geological stage. We observe systematic differences in extinction risk over time among marine genera classified according to their rarity. Geographic range played a primary role in determining extinction, and habitat breadth a secondary role, whereas local abundance had little effect. These results suggest that current reductions in geographic range size will lead to pronounced increases in long-term extinction risk even if local populations are relatively large at present.

Keywords: Phanerozoic; invertebrate; macroevolution; macroecology; conservation

1. INTRODUCTION

Rarity is widely used to assess extinction risk for conservation purposes [1] and has been implicated as a risk factor in past extinctions [2-7]. There are many different ways to be rare, but their relative influences on long-term extinction dynamics are poorly known. Species may be rare because they have small geographic ranges, narrow habitat tolerances, small populations or any combination thereof. When each of these measures of rarity is dichotomized, there are seven unique forms of rarity (e.g. small range, narrow habitat, small population; large range, narrow habitat, small population) [8]. Studies of contemporary risk reveal links between some aspects of rarity and population decline [9-12], but whether or not these declines will ultimately lead to extinction is, of course, uncertain. Furthermore, investigating the associations between rarity and extinction risk in extant species can easily become circular because some predictor variables, such as geographic range, are used in defining risk [13-15]. The fossil record provides the only opportunity to directly assess the relationships between rarity and extinction risk.

Previous studies have considered the effects of geographic range [3,5,6,16-22], and to a lesser extent, abundance [3,4,23-26] and habitat breadth [7,19,27], on extinction risk in the marine fossil record. However, no study has attempted to evaluate simultaneously the relative effects of rarity in all its different forms on long-term patterns of extinction risk. Because different aspects of rarity

*Author for correspondence: (paulharnik@nescent.org).

[†] Present address: National Evolutionary Synthesis Center, Durham, NC 27705, USA.

Electronic supplementary material is available at http://dx.doi.org/ 10.1098/rspb.2012.1902 or via http://rspb.royalsocietypublishing.org. often covary [3,28], and are measured in different units, it is impossible to assess their relative importance in extinction dynamics simply by comparing the results of univariate analyses, which make up the majority of the existing literature on extinction selectivity in the fossil record. The few multivariate studies that have been conducted illustrate clearly the impact that covariation can have on inferred causal relationships [3,6,29–31], but do not focus on rarity specifically. Moreover, all of these studies were limited in spatial, temporal and/or taxonomic scope.

In this study, we used the Paleobiology Database (http://paleodb.org) to investigate the associations between different forms of rarity and extinction risk for marine animals through the Phanerozoic (543 Ma-recent). Focusing on the three canonical aspects of rarity-geographic range size, habitat breadth and local abundance-that are the basis for assessing risk today [1], we quantified rarity and extinction risk within each geological stage for more than 6000 marine invertebrate genera representing more than 70 taxonomic classes and spanning a wide variety of functional groups. The complete dataset contains more than 13 000 stage-level observations of genera with associated ecological and extinction data and was used to examine the associations between rarity and extinction risk in a multivariate analytical framework. These analyses reveal systematic and persistent differences in extinction risk among different forms of rarity over the macroevolutionary history of marine animals.

2. MATERIAL AND METHODS

(a) Data download

The data used for our study were downloaded from the Paleobiology Database (http://paleodb.org) on 26 September 2010. The download included all occurrences entered by the Marine Invertebrate Working Group, excluding vertebrates and genera listed in quotation marks or qualified as '?', 'cf.' or 'aff.'. Vertebrates were not included in this study because of their relatively limited abundance and occurrence data. The following data associated with each occurrence were also downloaded: palaeolatitude, palaeolongitude, palaeoenvironment, primary lithology and taxon abundance. All occurrences were filtered to include only genera that were classified to a higher taxon (family, order and/or class) and which were assigned to one of the 74 geologic stages. The resulting dataset consisted of 301 904 occurrences in 42 467 collections for a total of 6491 genera. Overall, 72 classes are represented, with 27 classes having more than 1 per cent mean proportional diversity in the Palaeozoic, Mesozoic and/or Cenozoic Eras (see the electronic supplementary material, table S1).

(b) Measuring rarity and risk

Geographic range was measured as the area occupied by a genus relative to the maximum occupancy possible in an interval to account for temporal variation in the quality of the fossil record [3,21]. The globe was divided into 10 000 cells, each 3.6° longitude $\times 1.8^{\circ}$ latitude in size, and the range size was calculated as the number of cells in which a genus occurred relative to the total number of cells that contained fossils in the interval, using the palaeocoordinates of fossil occurrences. Although cell size varies with latitude, a previous study of the Paleobiology Database conducted at comparable spatial and temporal scales found little difference between results generated using equal area versus $5^{\circ} \times 5^{\circ}$ cells because high-latitude fossil occurrences are relatively limited overall, and tend to be concentrated in only a few regions during those intervals in which they occur [32]. Palaeocoordinates were derived from Scotese's palaeomap rotations, provided as part of the standard download protocol from the Paleobiology Database. Over the last 100 million years, palaeocoordinates are well constrained using magnetic sea floor anomaly data. Due to subduction of oceanic crust, greater uncertainty exists in palaeogeographic reconstructions, in particular palaeolongitude, deeper in time [33,34].

Habitat breadth was measured using three environmental variables that are important determinants of species distributions in the oceans today and which can be readily identified in the marine geologic record: relative water depth (above and/or below storm wave base), substrate (carbonate, siliciclastic and/or mixed) and latitude (tropical and/or extratropical). We defined 12 possible habitats based on all possible pairwise combinations of these three variables $(2 \times 3 \times 2)$ and counted up the number a genus occupied and divided that by the total number of habitats containing fossil occurrences in the interval to account for temporal variation in the preservation of environments. Storm wave base reflects the interactions between bathymetry, sedimentology and climate [35], and is best viewed as a relative, rather than absolute, measure of water depth [36]. We use storm wave base to delimit the broad-scale depth tolerance of genera and refrained from finer subdivisions of palaeowater depth because of greater uncertainty in defining these depth breaks across a diversity of sedimentary basins.

Local abundance for each genus was measured as its mean proportional abundance in those collections in which it occurred within an interval [23]; only collections containing 100 or more specimens were used for this calculation. Extinction risk was measured as the observed extinction or survival of a genus in an interval. Each measure of rarity was transformed to satisfy assumptions of normality using an arcsine square root transformation commonly applied to proportional data that include zero values and then scaled to zero mean and unit variance to allow relative effects on extinction risk to be assessed on a comparable scale.

These measures of rarity describe the summed geographic range, habitat breadth and local abundance of all congeneric species within each genus. Genus extinction similarly reflects the extinction of all the populations of its congeneric species. To the extent that many genera are monospecific, associations observed between rarity and extinction risk at the genus-level may be informative for understanding species-level dynamics.

(c) Modelling the relationships between rarity and extinction risk

We examined the association between rarity and extinction risk using two multivariate approaches. First, we assigned genera to one of the eight classes following the classification scheme of Rabinowitz [8]. We used the median values of abundance, range size and habitat breadth to delimit rare versus common taxa in each interval. We compared the odds of extinction (odds of extinction = q/(1-q), where q is the probability of extinction) of genera in the seven classes characterized by one or more aspects of rarity (e.g. small range, broad habitat, large population; small range, small habitat, large population) to the odds of extinction for genera in the eighth class that were considered common by all three rarity measures (i.e. large range, broad habitat, large population). Distinguishing rare from common genera using the median versus a different quantile is an arbitrary decision because each rarity measure is continuous. Because most genera exhibit values that are less than half of the maximum observed, the median provides a conservative estimate of differential risk. Secondly, we used multiple logistic regression to assess the associations between each rarity measure and extinction risk within a continuous framework. A multiple logistic regression model was fit to all intervals containing data for 50 or more genera.

There are strengths and limitations to each of these approaches. Analysis of rarity classes can reveal synergistic effects resulting from the interactions between different aspects of rarity but imposes discrete breaks on what are really continuous variables. Relative odds of extinction can also only be compared in intervals in which some common genera go extinct. Using continuous multiple logistic regression addresses these two limitations, allowing us to examine these associations without having to define arbitrary cut-offs between 'rare' and 'common' genera. However, in most intervals the sample size of genera with associated rarity data is too small to estimate interaction terms as well as additive effects: the median sample size of genera in an interval is 178, with a minimum of 51 and maximum of 800; moreover, stages containing many genera may still yield little statistical power if the extinction rate was either very high or very low and as a result few genera survived or went extinct.

To address the potential confounding effects of incomplete sampling [37–40], the number of occurrences was used as a measure of per-genus sampling probability in the continuous multiple logistic regression model. The number of genus occurrences in each interval was logarithmically transformed prior to analysis.



Rarity and long-term extinction risk P. G. Harnik et al. 3

Figure 1. The odds of extinction for genera characterized by different forms of rarity relative to genera considered common by all measures. Warmer colours indicate greater odds of extinction on a logarithmic scale. The solid horizontal lines denote equal odds of extinction between rare and common genera. Dashed lines and inset values denote the median relative odds of extinction for each rarity class over the last 500 million years. The odds ratio is presented for all intervals in which both rare and common genera were observed to go extinct, and cannot be calculated for the 18 shaded intervals (dark grey) in which no common genera were observed to go extinct.

3. RESULTS AND DISCUSSION

Our analysis revealed systematic differences in the odds of extinction among genera classified according to their rarity (figures 1 and 2). Geographic range had the strongest effect on extinction risk, with habitat breadth contributing secondarily. The median odds of extinction among genera with small ranges and narrow habitat breadth were six-times greater than for common genera (i.e. those with large populations, large ranges and broad habitat breadth). Because little extinction occurred among common genera in most stages, this finding indicates an approximately sixfold difference in the probability of extinction between these groups. Among genera with small ranges, broad habitat breadth was associated with a 30 per cent reduction in the median odds of extinction, yet these genera were still four times more likely to go extinct than common genera. In contrast, local abundance contributed remarkably little. The associations between rarity types and extinction risk were fairly consistent over geological time; although variation occurred among stages (e.g. weaker extinction selectivity during the Permian-Triassic and Cretaceous-Palaeogene mass extinctions), there is no evidence for systematic changes in the relationships between forms of rarity and extinction risk over the last 500 million years (figure 1).

The pattern of correlation between rarity measures and extinction risk is insensitive to analytical approach. A multiple logistic regression model in which the three aspects of rarity were treated as continuous additive predictors of extinction risk (see figure 3 and electronic supplementary material, table S2) yields comparable results to the analysis using discrete rarity classes. In this case, the odds ratio indicates the change in the odds of extinction associated with a one unit increase in the value of the predictor variable. Geographic range had a strong negative association with extinction, with more broadly distributed genera consistently at lower risk throughout the Phanerozoic. On average, habitat tolerance was also inversely associated with extinction risk, although this relationship was much more variable in the multiple logistic regression analysis. Local abundance showed little association with extinction risk, with genera that occurred at lower abundance at comparable risk to those that occurred at greater abundance.

Despite secular changes in our ability to resolve the geographic distributions of marine organisms in the fossil record (see §2*b*), this variation in the nature of the fossil record has no identifiable effect on the patterns of extinction selectivity presented here. There is no significant difference in the mean log odds of extinction according to the geographic range size between intervals before or after 100 Ma (*t*-test, p > 0.05), nor any secular trend in the association between the range size and extinction risk through the Phanerozoic (figure 3).

The associations between rarity and extinction risk revealed in the analysis that pooled data for all animal genera are also present when the analysis is conducted within more restricted taxonomic groupings. Bivalves, gastropods and brachiopods together comprise 72 per cent of the full dataset and each clade had sufficient data to examine the associations between the three continuous rarity measures and extinction risk. The multiple logistic regression model described above was fit separately to the data for each clade using stages in which more than 50 genera in the clade had associated rarity data. Results for each clade (figure 4) were comparable to those generated using the full dataset (figure 3), and there were no



Figure 2. The relative odds of extinction for the seven forms of rarity over the last 500 million years, with risk increasing from the bottom to top panels. Data are from figure 1, and are the number of time intervals characterized by different odds of extinction for genera in each rarity class relative to common genera. The solid black lines denote equal odds of extinction between rare and common genera. Dashed white lines denote the median relative odds of extinction for each rarity class over the last 500 million years. At the greatest risk of extinction are genera with both small ranges and narrow habitat breadth. Broader habitat breadth helps buffer narrowly distributed genera from extinction, whereas variation in the population size has negligible effect.

significant differences among clades in their distribution of log odds values for each rarity measure (p > 0.05 for each Kolmogorov–Smirnov test before Bonferroni correction for multiple comparisons).

Geographic range is the principal aspect of rarity associated with extinction risk in our multivariate analyses of marine invertebrates as a whole (figures 1-3), as well as multivariate analyses of diverse clades within the

Proc. R. Soc. B

overall dataset (bivalves, gastropods and brachiopods). However, abundance data are too limited for most phyla and classes in most intervals to fit the multivariate model. To investigate variability in the association between geographic range and extinction risk among clades, we fit a logistic regression model with range size as the sole predictor for seven animal phyla and for marine protists. The association between range size and extinction risk may bear the indirect contributions of habitat breadth and population size, but should be dominated by the direct effect of geographic range. All eight groups exhibit a consistent, inverse relationship between geographic range size and extinction risk (figure 5). The two groups that contain some pelagic taxa (Protozoa and Mollusca) exhibit positive log odds (i.e. a positive association between geographic range and extinction risk) in a limited number of intervals. These results indicate remarkably consistent patterns of extinction risk with respect to different forms of rarity across both time and higher taxa.

Could the persistent correlation between rarity and extinction risk across geological time and higher taxa result from sampling biases rather than biological processes? Incomplete sampling could generate apparent differences in extinction risk among rare versus common genera by spuriously shortening the durations of rare taxa [29,37-40]. We examined this potential bias in two ways and found our results to be robust to these alternative treatments of the data. First, we added occurrence frequency as a measure of per-interval sampling probability for each genus to our multiple logistic regression model (figure 3) and found little change in the estimated associations between the three aspects of rarity and extinction risk. Parameter estimates were strongly correlated between the two models (abundance: Spearman $\rho=0.95, \ p<0.0001;$ geographic range: Spearman $\rho = 0.67, \, p < 0.0001;$ habitat breadth: Spearman $\rho = 0.95,$ p < 0.0001) and there was equivocal support for the model that included this additional parameter (electronic supplementary material, figure S1). Furthermore, models that included geographic range or all three rarity measures in addition to occurrence frequency tended to have much greater support than a model of extinction risk based on number of occurrences alone (electronic supplementary material, figure S2). Secondly, we excluded genera observed in only one interval (singletons) from our multiple logistic regression analysis and found that this also had little effect on the estimated associations between rarity and extinction risk. Parameters estimated using the data for all genera were strongly correlated with those estimated using the data excluding singleton genera (abundance: Spearman $\rho = 0.86$, p < 0.0001; geographic range: Spearman $\rho = 0.84$, p < 0.0001; habitat breadth: Spearman $\rho = 0.90$, p < 0.0001). These two approaches are conservative; the frequency of singletons and the number of occurrences of genera in an interval result from the interplay between rarity and sampling yet are assumed in these two treatments to be exclusively sampling artefacts. These results are also congruent with several previous studies [3,5,21,23], which have shown that the associations between individual aspects of rarity and extinction risk could not be attributed simply to sampling biases.

More broadly, there is little reason to expect that sampling biases can explain the long-term associations



Figure 3. The relationships between extinction risk and abundance, geographic range and habitat breadth over the last 500 million years estimated using a multiple logistic regression model containing the three continuous predictor variables. The solid line indicates no association between the predictor variable and extinction risk. Negative values indicate an inverse association between the predictor risk. Histograms for each aspect of rarity present the number of time intervals characterized by different log odds of extinction. The associations between each aspect of rarity and extinction risk in this continuous analysis are comparable with those observed in the analysis of discrete rarity classes.

between rarity and extinction risk that we observe here. First, we are not attempting to evaluate absolute changes in a given rarity metric over time; rather, we are asking whether the values of rarity metrics for genera within a given time interval are associated with either extinction or survival at the end of that same interval. If many rare taxa in one interval (t) appeared to go extinct, but in fact survived into the next interval (t+1), this could enhance the apparent extinction selectivity in interval t. However, these taxa should then be added to the list of unsampled victims for some subsequent time interval (likely t+1) and so result in an underestimation of extinction selectivity in that interval due to the failure to sample many rare victims. Thus, while variation in sampling intensity could create artificially strong or weak selectivity within any given interval, it could not produce the consistent interval-after-interval selectivity observed in this study unless most of the unsampled survivors persist to the present day, which is unrealistic.

The strong association between geographic range and extinction risk documented here has been observed previously in studies focused solely on geographic range

(e.g. [18,21]), and in the few multivariate analyses conducted at primarily finer spatial, temporal and taxonomic scales [3,6,16,29-31]. Among early Cenozoic bivalves in North America, for example, geographic range was the only biological factor consistently associated with species duration after accounting for covariation with local abundance and body size [3]. Similarly, among Cenozoic mollusks in South America [31] and New Zealand [29], geographic range was strongly and consistently associated with species duration even after factoring out the effects of life habit and other biological characteristics. At the global scale, geographic range was significantly associated with survivorship among bivalve genera during the Cretaceous-Palaeogene mass extinction [6], and among skeletonized marine invertebrate genera through the Phanerozoic [16], and in both of these studies the association between geographic range and survivorship remained after accounting for variation in species richness. These previous studies corroborate the strong effect of geographic range on extinction risk, yet the current study is the first to establish the importance of range size relative to all other forms of rarity.



Figure 4. The log odds of extinction for gastropod, bivalve and brachiopod genera over geologic time according to their abundance, geographic range and habitat breadth. Values are the median (circles) and interquartile range of log odds values for geological stages, generated by fitting a continuous multiple logistic regression model separately to data for each clade in stages for which more than 50 genera in the clade have associated rarity data. The solid vertical line indicates no association between the predictor variable and extinction risk. Each group shows selectivity patterns consistent with the pooled analysis and there are no significant differences between clades in their distributions of log odds values for each predictor variable.

One widely accepted explanation for the pervasive association between geographic range size and extinction risk is that large ranges buffer taxa from biotic and abiotic stresses affecting more limited geographic areas. The critical factor is the size of a taxon's range and only secondarily how that range is distributed across habitats (this study, [18,19]). Despite the importance of geographic range in macroevolution, relatively little is known about the factors that generate variation in geographic range size over deep time. To what extent can the substantial variation in range size observed across taxa be explained by differences in fecundity, dispersal, competition and environmental preference [3,7,30,41]? Have these same attributes, in conjunction with the expansion and contraction of environments over geologic time, given rise to the



Figure 5. The log odds of extinction for genera belonging to different major marine clades over the Phanerozoic according to their geographic range. Values are the median (circles) and range (minimum to maximum) of log odds values, generated by fitting a continuous logistic regression model separately to the data for each clade using stages in which more than 50 genera in the clade had associated geographic range data. The solid vertical line indicates no association between the geographic range and extinction risk. All seven animal phyla and marine protists exhibit a consistent inverse relationship between geographic range size and extinction risk, with only two clades which contain some pelagic taxa (Protozoa and Mollusca) exhibiting positive log odds values in a limited number of intervals.

long-term changes in the range size observed over the histories of individual marine taxa [42-44]? And what role, if any, has extinction selectivity played in shaping long-term changes in the average range size of marine faunas (e.g. [45, fig. 1c])?

The lack of association between local abundance and extinction risk over geologic time is surprising given the extensive literature documenting effects of population size on extinction risk among extant species [11,12,46-48]. Yet our results are in accord with other analyses of the fossil record that have found no association between abundance and extinction risk [24], or a negative association that was due entirely to covariation between abundance and geographic range size [3]. More broadly, positive [49], negative [4,50] and non-monotonic [23] relationships have all been reported in palaeontological studies. This degree of heterogeneity contrasts markedly with the consistent negative association between geographic range size and extinction risk, and strongly suggests that taxa observed at low abundance in the fossil record had population sizes that were considerably greater than the minimum size below which the effects of demographic stochasticity become critical [51,52] and/or possessed traits that allowed them to counteract the problems of reproduction and recruitment at low densities. Actualistic studies comparing the abundance of species in contemporary living communities with their associated time-averaged death assemblages [53] may help to identify threshold population sizes below which

species are so rare they are unlikely to be preserved and subsequently sampled.

4. CONCLUSIONS

Short- and long-term monitoring data are lacking for most marine species today [54-56]. As a result, assessments of extinction risk for living marine animal species are typically made on the basis of qualitative or semi-quantitative measures of overall abundance and geographic range size [1]. The fossil record shows that variation in geographic range size has the dominant effect on extinction risk over long timescales, and that this association is due to the buffering effects of range size and not the greater habitat tolerance or larger local population size of broadly distributed taxa. This pattern of extinction selectivity through the Phanerozoic marine fossil record fingerprints regional-scale environmental perturbations and not demographic stochasticity as the principal driver of past extinctions, thereby highlighting the utility of the fossil record for identifying the traits that lead to elevated risk.

Although extinction occurs when all the populations of a taxon decline to zero, the primary biological predictor of this process over the last 500 million years has been geographic range size followed by habitat breadth, but not local abundance. These results are robust to varying analytical approach and to alternative treatments of the data that address the potential biasing effects of incomplete sampling. Moreover, this pattern has held consistent throughout the history of marine animal life and within diverse animal phyla and classes. Therefore, it appears unlikely to reflect either the circumstances of any particular time in Earth history or clade-specific ecological or physiological traits. Taken together these results suggest that contemporary reductions in the range size observed in many groups [57] will be accompanied by a pronounced increase in long-term extinction risk even if population sizes remain relatively large at present.

We thank the researchers who collected these data and the members of the Paleobiology Database who compiled them, M. Elmore, M. Foote, J. A. Meachen, C. R. McClain, J. L. McGuire, and V. L. Roth for discussion, and D. L. Rabosky, P. J. Wagner and two anonymous reviewers for thoughtful comments on the present manuscript. This work was partially funded by NESCent (NSF grant EF-0905606) and DFG grant KI 806/7-1. This article is Paleobiology Database Publication no. 170. Data deposited at Dryad: doi:10.5061/dryad.0mq69.

REFERENCES

- 1 IUCN 2001 *Red list categories and criteria*, version 3.1. IUCN Species Survival Commission. Gland, Switzerland: IUCN.
- 2 McKinney, M. L. 1997 Extinction vulnerability and selectivity: combining ecological and paleontological views. *Annu. Rev. Ecol. Syst.* 28, 495–516. (doi:10. 1146/annurev.ecolsys.28.1.495)
- 3 Harnik, P. G. 2011 Direct and indirect effects of biological factors on extinction risk in fossil bivalves. *Proc. Natl Acad. Sci. USA* 108, 13 594–13 599. (doi:10.1073/pnas. 1100572108)
- 4 Payne, J. L., Truebe, S., Nützel, A. & Chang, E. T. 2011 Local and global abundance associated with extinction

risk in late Paleozoic and early Mesozoic gastropods. *Paleobiology* **37**, 616–632. (doi:10.1666/10037.1)

- 5 Kiessling, W. & Aberhan, M. 2007 Geographical distribution and extinction risk: lessons from Triassic– Jurassic marine benthic organisms. *J. Biogeogr.* 34, 1473–1489. (doi:10.1111/j.1365-2699.2007.01709.x)
- 6 Jablonski, D. 2008 Extinction and the spatial dynamics of biodiversity. *Proc. Natl Acad. Sci. USA* **105**, 11528– 11535. (doi:10.1073/pnas.0801919105)
- 7 Heim, N. A. & Peters, S. E. 2011 Regional environmental breadth predicts geographic range and longevity in fossil marine genera. *PLoS ONE* 6, e18946. (doi:10.1371/ journal.pone.0018946)
- 8 Rabinowitz, D. 1981 Seven forms of rarity. In *The biological aspects of rare plant conservation* (ed. J. Synge), pp. 205–217. Chichester, UK: Wiley.
- 9 Fisher, D. O. & Owens, I. P. F. 2004 The comparative method in conservation biology. *Trends Ecol. Evol.* 19, 391–398. (doi:10.1016/j.tree.2004.05.004)
- 10 Lee, T. M. & Jetz, W. 2010 Unravelling the structure of species extinction risk for predictive conservation science. *Proc. R. Soc. B* 278, 1329–1338. (doi:10.1098/rspb. 2010.1877)
- 11 Davidson, A. D., Hamilton, M. J., Boyer, A. G., Brown, J. H. & Ceballos, G. 2009 Multiple ecological pathways to extinction in mammals. *Proc. Natl Acad. Sci. USA* **106**, 10702–10705. (doi:10.1073/pnas.0901956106)
- 12 Cardillo, M., Mace, G. M., Gittleman, J. L., Jones, K. E., Bielby, J. & Purvis, A. 2008 The predictability of extinction: biological and external correlates of decline in mammals. *Proc. R. Soc. B* 275, 1441–1448. (doi:10. 1098/rspb.2008.0179)
- Purvis, A., Cardillo, M., Grenyer, R. & Collen, B. 2005 Correlates of extinction risk: phylogeny, biology, threat and scale. In *Phylogeny and conservation* (eds A. Purvis, J. L. Gittleman & T. Brooks), pp. 295–316. Cambridge, UK: Cambridge University Press.
- 14 Harcourt, A. H. 2005 Problems of studying extinction risks. *Science* 310, 1276–1277. (doi:10.1126/science. 310.5752.1276c)
- 15 Purvis, A., Gittleman, J. L., Cowlishaw, G. & Mace, G. M. 2000 Predicting extinction risk in declining species. *Proc. R. Soc. Lond. B* 267, 1947–1952. (doi:10. 1098/rspb.2000.1234)
- 16 Payne, J. L. & Finnegan, S. 2007 The effect of geographic range on extinction risk during background and mass extinction. *Proc. Natl Acad. Sci. USA* 104, 10506-10511. (doi:10.1073/pnas.0701257104)
- Hopkins, M. J. 2011 How species longevity, intraspecific morphological variation, and geographic range size are related: a comparison using Late Cambrian trilobites. *Evolution* 65, 3253–3273. (doi:10.1111/j.1558-5646. 2011.01379.x)
- 18 Powell, M. G. 2007 Geographic range and genus longevity of late Paleozoic brachiopods. *Paleobiology* 33, 530–546. (doi:10.1666/07011.1)
- 19 Jablonski, D. & Raup, D. M. 1995 Selectivity of end-Cretaceous marine bivalve extinctions. *Science* 268, 389-391. (doi:10.1126/science.11536722)
- 20 Liow, L. H. 2007 Does versatility as measured by geographic range, bathymetric range and morphological variability contribute to taxon longevity? *Glob. Ecol. Biogeogr.* **16**, 117–128. (doi:10.1111/j.1466-8238. 2006.00269.x)
- 21 Foote, M., Crampton, J. S., Beu, A. G. & Cooper, R. A. 2008 On the bidirectional relationship between geographic range and taxonomic duration. *Paleobiology* 34, 421–433. (doi:10.1666/08023.1)
- 22 Powell, M. G. & MacGregor, J. 2011 A geographic test of species selection using planktonic foraminifera during

8 P. G. Harnik et al. Rarity and long-term extinction risk

the Cretaceous/Paleogene mass extinction. *Paleobiology* **37**, 426–437. (doi:10.1666/10010.1)

- 23 Simpson, C. & Harnik, P. G. 2009 Assessing the role of abundance in marine bivalve extinction over the post-Paleozoic. *Paleobiology* 35, 631–647. (doi:10.1666/ 0094-8373-35.4.631)
- 24 Lockwood, R. 2003 Abundance not linked to survival across the end-Cretaceous mass extinction: patterns in North American bivalves. *Proc. Natl Acad. Sci. USA* 100, 2478–2482. (doi:10.1073/pnas.0535132100)
- 25 McClure, M. & Bohonak, A. J. 1995 Non-selectivity in extinction of bivalves in the Late Cretaceous of the Atlantic and Gulf Coastal Plain of North America. *J. Evol. Biol.* 8, 779–794. (doi:10.1046/j.1420-9101. 1995.8060779.x)
- 26 Hotton, C. L. 2002 Palynology of the Cretaceous– Tertiary boundary in central Montana: evidence for extraterrestrial impact as a cause of the terminal Cretaceous extinctions. *Geol. Soc. Am. Special Paper* 361, 473–501. (doi:10.1130/0-8137-2361-2.473)
- 27 Kammer, T. W., Baumiller, T. K. & Ausich, W. I. 1998 Evolutionary significance of differential species longevity on Osagean–Meramecian (Mississippian) crinoid clades. *Paleobiology* 24, 155–176.
- 28 Gaston, K. J. 1994 Rarity. London, UK: Chapman & Hall.
- 29 Crampton, J. S., Cooper, R. A., Beu, A. G., Foote, M. & Marshall, B. A. 2010 Biotic influences on species duration: interactions between traits in marine molluscs. *Paleobiology* 36, 204–223. (doi:10.1666/09010.1)
- 30 Jablonski, D. & Hunt, G. 2006 Larval ecology, geographic range, and species survivorship in Cretaceous mollusks: organismic versus species-level explanations. *Am. Nat.* 168, 556–564. (doi:10.1086/507994)
- 31 Rivadeneira, M. M. & Marquet, P. A. 2007 Selective extinction of late Neogene bivalves on the temperate Pacific coast of South America. *Paleobiology* 33, 455–468. (doi:10.1666/06042.1)
- 32 Miller, A. I., Aberhan, M., Buick, D. P., Bulinski, K. V., Ferguson, C. A., Hendy, A. J. W. & Kiessling, W. 2009 Phanerozoic trends in the global geographic disparity of marine biotas. *Paleobiology* 35, 612–630. (doi:10.1666/ 0094-8373-35.4.612)
- 33 Schmachtenberg, W. F. 2011 Paleolongitudinal estimates for paleocontinents derived from interplate distances based on Late Ordovician bivalves. *Paleobiology* 37, 438–444. (doi:10.1666/09015.1)
- 34 Scotese, C. R., Bambach, R. K., Barton, C., Van der Voo, R. & Ziegler, A. M. 1979 Paleozoic base maps. *f. Geol.* 87, 217–277. (doi:10.1086/628416)
- 35 Immenhauser, A. 2009 Estimating palaeo-water depth from the physical rock record. *Earth Sci. Rev.* **96**, 107–139. (doi:10.1016/j.earscirev.2009.06.003)
- 36 Peters, S. E. & Loss, D. P. 2012 Storm and fair-weather wave base: a relevant distinction? *Geology* 40, 511–514. (doi:10.1130/g32791.1)
- 37 Russell, M. P. & Lindberg, D. R. 1988 Real and random patterns associated with molluscan spatial and temporal distributions. *Paleobiology* 14, 322–330.
- 38 Holland, S. M. & Patzkowsky, M. E. 2002 Stratigraphic variation in the timing of first and last occurrences. *Palaios* 17, 134–146. (doi:10.1669/0883-1351(2002) 017<0134:SVITTO>2.0.CO;2)
- 39 Meldahl, K. H. 1990 Sampling, species abundance, and the stratigraphic signature of mass extinction: a test using Holocene tidal flat mollusks. *Geology* 18, 890–893. (doi:10.1130/0091-7613(1990)018<0890:SS AATS>2.3.CO;2)

- 40 Signor III, P. W & Lipps, J. H. 1982 Sampling bias, gradual extinction patterns, and catastrophes in the fossil record. *Geol. Soc. Am. Special Paper* 190, 291–296.
- 41 Roy, K., Hunt, G., Jablonski, D., Krug, A. Z. & Valentine, J. W. 2009 A macroevolutionary perspective on species range limits. *Proc. R. Soc. B* 276, 1485–1493. (doi:10. 1098/rspb.2008.1232)
- 42 Foote, M. 2007 Symmetric waxing and waning of marine invertebrate genera. *Paleobiology* 33, 517–529. (doi:10. 1666/06084.1)
- 43 Foote, M., Crampton, J. S., Beu, A. G., Marshall, B. A., Cooper, R. A., Maxwell, P. A. & Matcham, I. 2007 Rise and fall of species occupancy in Cenozoic fossil mollusks. *Science* 318, 1131–1134. (doi:10.1126/science.1146303)
- 44 Liow, L. H. & Stenseth, N. C. 2007 The rise and fall of species: implications for macroevolutionary and macroecological studies. *Proc. R. Soc. B* 274, 2745–2752. (doi:10.1098/rspb.2007.1006)
- 45 Miller, A. I. & Foote, M. 2009 Epicontinental seas versus open-ocean settings: the kinetics of mass extinction and origination. *Science* **326**, 1106–1109. (doi:10.1126/ science.1180061)
- 46 Boyce, M. S. 1992 Population viability analysis. Annu. Rev. Ecol. Syst 23, 481–506. (doi:10.1146/annurev. ecolsys.23.1.481)
- 47 Brook, B. W., O'Grady, J. J., Chapman, A. P., Burgman, M. A., Akcakaya, H. R. & Frankham, R. 2000 Predictive accuracy of population viability analysis in conservation biology. *Nature* 404, 385–387. (doi:10.1038/35006050)
- 48 O'Grady, J. J., Reed, D. H., Brook, B. W. & Frankham, R. 2004 What are the best correlates of predicted extinction risk? *Biol. Conserv.* 118, 513–520. (doi:10.1016/j,biocon. 2003.10.002)
- 49 Layou, K. M. 2007 A quantitative null model of additive diversity partitioning: examining the response of beta diversity to extinction. *Paleobiology* **33**, 116–124. (doi:10.1666/06025.1)
- 50 Zaffos, A. & Holland, S. M. 2012 Abundance and extinction in Ordovician–Silurian brachiopods, Cincinnati Arch, Kentucky and Ohio. *Paleobiology* 38, 278–291. (doi:10.5061/dryad.rj28ck8g)
- 51 Lande, R. 1993 Risks of population extinction from demographic and environmental stochasticity and random catastrophes. *Am. Nat.* 142, 911–927. (doi:10. 1086/285580)
- 52 Brook, B. W., Sodhi, N. S. & Bradshaw, C. J. A. 2008 Synergies among extinction drivers under global change. *Trends Ecol. Evol.* **23**, 453–460. (doi:10.1016/j. tree.2008.03.011)
- 53 Kidwell, S. M. 2001 Preservation of species abundance in marine death assemblages. *Science* 294, 1091–1094. (doi:10.1126/science.1064539)
- 54 Harnik, P. G *et al.* 2012 Extinctions in ancient and modern seas. *Trends Ecol. Evol.* 27, 608–617. (doi:10. 1016/j.tree.2012.07.010)
- 55 McClenachan, L., Cooper, A. B., Carpenter, K. E. & Dulvy, N. K. 2012 Extinction risk and bottlenecks in the conservation of charismatic marine species. *Conserv. Lett.* 5, 73–80. (doi:10.1111/j.1755-263X. 2011.00206.x)
- 56 Regnier, C., Fontaine, B. & Bouchet, P. 2009 Not knowing, not recording, not listing: numerous unnoticed mollusk extinctions. *Conserv. Biol.* 23, 1214–1221. (doi:10.1111/j.1523-1739.2009.01245.x)
- 57 Ceballos, G. & Ehrlich, P. R. 2002 Mammal population losses and the extinction crisis. *Science* 296, 904–907. (doi:10.1126/science.1069349)